RSC Advances RSC Publishing

PAPER

View Article Online

Efficient removal of cobalt from aqueous solution using β-cyclodextrin modified graphene oxide

Cite this: RSC Advances, 2013, 3, 9514

Wencheng Song,^{ab} Jun Hu,^b Ying Zhao,^b Dadong Shao*^b and Jiaxing Li^b

β-Cyclodextrin modified graphene oxide (denoted as β-CD–GO) was synthesized and applied as a sorbent to remove Co(II) from aqueous solution. The synthesized β-CD–GO was characterized by scanning electron microscopy and energy dispersive X-ray spectroscopy (SEM/EDX), Fourier transform infrared (FT-IR) spectroscopy, Raman spectroscopy and X-ray photoelectron spectroscopy (XPS) in detail, and the results of the characterization indicated that β-CD was successfully grafted onto the surface of GO. The effects of pH, ionic strength, competing ions and solid content on the sorption of Co(II) onto the β-CD–GO nanocomposites were investigated. The results indicated that Co(II) sorption on β-CD–GO was strongly dependent on pH and independent of ionic strength under the experimental uncertainties. The sorption isotherms of Co(II) on β-CD–GO could be described well by the Langmuir model. The thermodynamic data (ΔH^0 , ΔS^0 and ΔG^0) calculated from the temperature-dependent sorption isotherms suggest that the sorption of Co(II) on β-CD–GO is a spontaneous and endothermic process. The maximum sorption capacity of Co(II) on β-CD–GO (72.4 mg g⁻¹) at pH = 6.0 ± 0.1 and T = 303 K was much higher than that of Co(II) on GO (47.39 mg g⁻¹) and other sorbents. The abundant oxygen-containing functional groups of modified β-CD on the GO surface can improve the sorption capacity of β-CD–GO for Co(II). The results indicate that β-CD–GO is a promising material for the removal and enrichment of Co(III) from wastewater.

Received 24th March 2013, Accepted 8th April 2013

DOI: 10.1039/c3ra41434e

www.rsc.org/advances

1 Introduction

Radionuclides 60Co and 58Co are abundant in the nuclear waste released from pressurized water nuclear reactors, and ⁶⁰Co is also widely used in research and medical applications. ¹ The release of Co(II) into the environment could result in defects, diarrhea, low blood pressure, lung irritation or paralysis, and may also cause mutations (genetic changes) in living cells.2 Therefore, the removal of Co(II) from aqueous solution is important for public health. Over the past decades, various methods such as precipitation, ion-exchange, and sorption have been employed to remove Co(II) from large volumes of aqueous solution.3-5 Sorption technology is widely regarded as one of the most effective choices for the removal of heavy metal ions from aqueous solution because it is simple and cost-effective. 6,7 However, most of the common sorbents, such as clay minerals (i.e. maximum sorption capacity ~ 13.2 mg g⁻¹ Co(II) on sludge⁸), metallic oxides (i.e. maximum \sim 28.3 mg g⁻¹ Co(II) on γ -Fe₂O₃⁹) and carbon materials (*i.e.* maximum $\sim 4.1 \text{ mg g}^{-1} \text{ Co(II)}$ on poly(acrylonitrile-g-starch)¹⁰ and $\sim 12.4 \text{ mg g}^{-1} \text{ Co(II)}$ on bis(imidazole)-modified glycidyl methacrylate copolymer¹¹), usually suffer from either low

Graphene has a large specific surface area (theoretical value is 2620 m² g⁻¹),¹² which gives it the potential for application in the removal of pollutants in environmental pollution management.^{13–15} However, the strong aggregation of graphene nanosheets, the absence of effective methods of dispersing graphene in aqueous solution¹⁶ and the presence of less effective functional groups for sorption applications on its surface greatly restrict the application of graphene in environmental management. Therefore, it is necessary to modify the surface of graphene to enhance its sorption capacity and dispersion properties in aqueous solution.

Cyclodextrins (CDs) are cyclic oligosaccharides with a characteristic truncated cone shape which forms a well-defined cylindrical cavity. 17 β -CD, the most common CD, is characterized by a hydrophilic exterior and a hydrophobic interior, 18 and has the tendency to form inclusion complexes with contaminants such as organic pollutants and heavy metal ions. 19,20 Therefore, the modification of β -CD on the graphene oxide (GO) surface can effectively enhance the sorption capacity of GO.

In this study, β -CD modified GO (denoted as β -CD-GO) was synthesized by an *in situ* polymerization technique and

E-mail: shaodadong@126.com; Fax: +86-551-5591310; Tel: +86-551-5592788

sorption capacities or sorption efficiencies in the removal of heavy metal ions from aqueous solution. Therefore, the design and development of special sorbents with high sorption capacities for contaminants is critical for pollution management and related applications.

^aSchool of Nuclear Science and Technology, University Science and Technology of China, 230026 Hefei, P. R. China

 $[^]b$ Key Laboratory of Novel Thin Film Solar Cells, Institute of Plasma Physics, Chinese Academy of Sciences, P.O. Box 1126, 230031 Hefei, P. R. China.

RSC Advances Paper

characterized by scanning electron microscopy–energy dispersive X-ray spectroscopy (SEM/EDX), Fourier transform infrared spectroscopy (FT-IR), Raman spectroscopy and X-ray photoelectron spectroscopy (XPS). The batch sorption technique was adopted to evaluate the effect of different experimental conditions on the sorption performance of $\beta\text{-CD-GO}$ in the removal of Co(II) from aqueous solution.

2 Experimental section

2.1. Preparation of GO

GO was synthesized from flake graphite by a modified Hummers method. 21,22 Under mechanical stirring and in an ice–water bath, 98% $\rm H_2SO_4$ and $\rm KMnO_4$ were slowly added into a three-necked flask containing graphite powder and NaNO_3, and then the mixture reacted at room temperature for 5 days. After dilution with Milli-Q water at 98 $^{\circ}\rm C$, and cooling to 60 $^{\circ}\rm C$, $\rm H_2O_2$ (30%) was added to the mixture to reduce the residual KMnO_4. Finally, the solid GO was obtained after the black deposit was filtered and thoroughly washed with Milli-Q water and alcohol.

2.2. Preparation of β-CD-GO

β-CD–GO was synthesized according to previous work by Guo. 23 After a 50 mL mixture of GO (0.05 g L $^{-1}$) and β -CD (18 g L $^{-1}$) was adjusted to pH 11 by ammonia solution, the solution was placed in a water bath (60 °C) and 2 mL hydrazine solution was added to initiate the modification reaction. After reaction at 60 °C for 4 h, a stable black dispersion was obtained. After filtration with a nylon membrane (0.22 μm) and being washed thoroughly with Milli-Q water, the dispersion was dried at 60 °C for 24 h in a vacuum oven. Thus β -CD–GO was obtained and used in the following experiments.

2.3. Characterization

The synthesized β -CD–GO and GO were characterized by SEM/EDX, FT-IR, Raman spectroscopy and XPS. The SEM/EDX investigations were carried out on a JSM-6320F field emission scanning electron microscopy (FESEM) instrument coupled with an EDX unit (INCA). FT-IR spectroscopy was carried out on Perki Elmer 100 spectrometer in a KBr pellet. The Raman spectrum was carried out on a LabRam HR Raman spectrometer at an excitation wavelength of 514.5 nm by an Ar⁺ laser. The XPS measurements were performed with an ESCALab220i-XL surface microanalysis system (VG Scientific) equipped with an Al K α ($h\lambda$ = 1486.6 eV) source.

2.4. Sorption experiments

The sorption experiments were carried out using the batch sorption technique in polyethylene centrifuge tubes. The sorbents, NaCl and Co(II) were added to achieve the desired concentrations of the different components. The pH was adjusted to the desired values by adding negligible amounts of 0.1 or 0.01 M HCl or NaOH. The test tubes were shaken for 24 h to achieve sorption equilibrium and then centrifuged for 20 min at 8000 r min $^{-1}$. The concentration of Co(II) was analyzed with an atomic absorption spectrophotometer. All of the

experimental data were the averages of triplicate determinations. The relative errors of the data were about 5%. The amount of Co(II) adsorbed on the β -CD–GO or GO surface was calculated from the difference between the initial concentration and the final concentration. The sorption percentage and the distribution coefficient (K_d) were calculated from the following equations:

$$K_{\rm d} = \frac{C_0 - C_{\rm e}}{C_{\rm e}} \frac{V}{m} \tag{1}$$

$$Sorption\% = \frac{C_0 - C_e}{C_0} \times 100$$
 (2)

where C_0 (mg L⁻¹) is the initial concentration, C_e (mg L⁻¹) is the equilibrium concentration, m (g) is the mass of β-CD-GO and V (L) is the volume of the suspension.

3 Results and discussion

3.1. Characterization of β-CD-GO and GO

The microstructures of GO and β -CD–GO can be observed by SEM imaging (Fig. 1). Differences in the surface morphology between GO and β -CD–GO can be observed. Fig. 1a shows that the surface of GO is smooth and there are few wrinkles, whereas the surface of β -CD–GO is rougher and features more wrinkles (Fig. 1b), which is important for preventing the aggregation of GO. Additionally, there are many packed nanoparticles which suggests that β -CD molecules have been successfully attached to the GO surface. This was also confirmed by EDX characterization. The atomic percentages of used graphite, prepared GO and β -CD–GO were 99.33% C and 0.67% O, 88.96% C and 11.03% O, and 94.49% C and 5.52% O respectively (Table 1).

The FT-IR spectra of GO, β -CD and β -CD–GO are shown in Fig. 2. The peak at $\sim\!3400~\text{cm}^{-1}$ is due to the O–H stretching vibration of hydroxyl groups on the sorbent surface. In the GO spectrum, a C–O group ($\sim\!1220~\text{cm}^{-1}$ and $\sim\!1100~\text{cm}^{-1}$), C=O group ($\sim\!1730~\text{cm}^{-1}$) and C=C bond ($\sim\!1620~\text{cm}^{-1}$) can be seen, which indicates that large amounts of oxygen-containing functional groups exist on the GO surface. In the β -CD–GO spectrum, a significant increase in band intensity is visible at $\sim\!1030~\text{cm}^{-1}$ and $\sim\!2920~\text{cm}^{-1}$, which can be attributed to the surface modified β -CD. It further confirms the successful modification of GO with β -CD.

In the Raman spectra (Fig. 3), the G band at <1580 cm⁻¹ is associated with the vibration of sp² carbon atoms in a

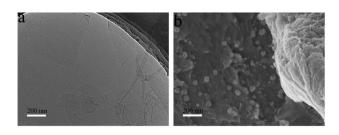


Fig. 1 SEM images of (a) GO and (b) β-CD-GO.

Paper

Table 1 Elemental composition of used graphite, GO and $\beta\text{-CD--GO},$ obtained by EDX

	C (at%)	O (at%)
Used graphite	99.33	0.67
GO	88.96	11.03
β-CD-GO	94.49	5.52

graphitic 2D hexagonal lattice, and the D band at <1350 cm $^{-1}$ is related to the vibrations of the sp 3 carbon atoms in defects and disorders. The value of the $I_{\rm G}/I_{\rm D}$ ratio can be used to study the relative extent of GO structural graphite. The $I_{\rm G}/I_{\rm D}$ ratios are 0.93 for GO and 0.87 for β -CD–GO. The $I_{\rm G}/I_{\rm D}$ ratio of GO decreases markedly after modification with β -CD, which could be due to the effect of the modified β -CD on the GO surface, because the modified β -CD on the GO surface would exist in a disordered form.

XPS (Fig. 4) is an effective technique for the analysis and identification of surface groups. The O 1s spectrum of β-CD-GO can be deconvoluted into three peaks at 531.4 ± 0.2 , 532.7 \pm 0.2, and 533.7 \pm 0.2 eV, which correspond to -COOH, >C=O and -C-OH species, respectively. The quantitative analysis of the O 1s spectrum of GO gives peak fractions of 46.7% for -COOH, 44.9% for >C=O and 8.38% for -C-OH, whereas those of the O 1s spectrum of β-CD-GO are 35.8% for -COOH, 28.2% for >C=O and 36.0% for -C-OH (Table 2). One can see that the peak fractions of -COOH and >C=O decrease, whereas the peak fraction of -C-OH increases after GO was modified with β -CD. This is reasonable because β -CD is a cyclic oligosaccharide which contains many hydroxyl groups. The modified β-CD molecules on GO increase the peak fraction of -C-OH and thereby decrease the other peak fractions.24,25

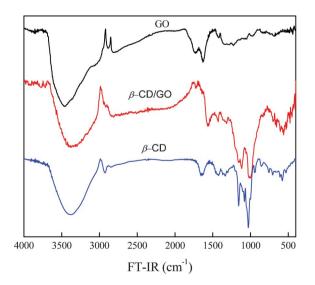


Fig. 2 FT-IR spectra of GO, β -CD and β -CD-GO.

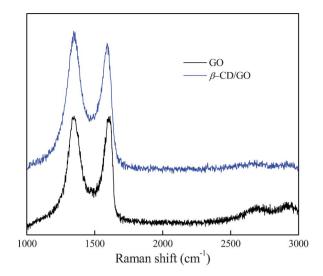


Fig. 3 Raman spectra of GO and β-CD-GO

3.2. Effect of pH and ionic strength

The pH of aqueous solution (Fig. 5A) is an important factor which affects the sorption of Co(II) because a change in the solution pH can influence the surface properties of the sorbents and the hydrolysis, precipitation and complexation of Co(II). The sorption of Co(II) on β -CD–GO as a function of pH in 0.1, 0.01, and 0.001 M NaCl solutions is shown in Fig. 5A. It can seen that the sorption of Co(II) increases slowly at pH < 6, then quickly at pH 6–9, and maintains high level of sorption at pH > 9. This strong pH-dependent sorption suggests that the sorption of Co(II) is mainly dominated by surface complexation. 27

The relative distribution of Co(II) species in aqueous solution can be calculated from the hydrolysis constants of Co(II) (log $K_1 = -9.6$, log $K_2 = -9.2$, and log $K_3 = -12.7$).²⁸ The distribution of different Co(II) species in aqueous solution as a function of solution pH is shown in Fig. 5B. Co(II) exists as the

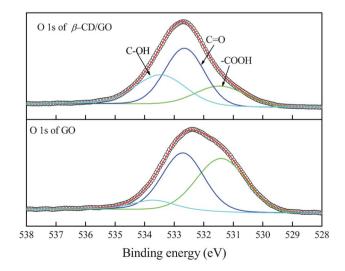


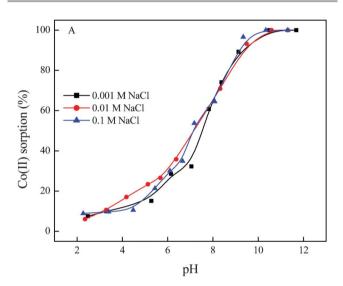
Fig. 4 XPS spectra of GO and β -CD-GO.

Table 2 Curve fitting results of the XPS O 1s spectra

	Peak	BE ^a (eV)	FWHM ^b (eV)	%
GO	-COOH	531.40	1.98	46.7
	>C $=$ O	532.70	1.67	44.9
	-C-OH	533.70	1.67	8.38
β-CD-GO	-COOH	531.44	2.02	35.8
	>C $=$ O	532.65	1.58	28.2
	-C-OH	533.66	2.03	36.0

^a Binding energy. ^b Full width at half-maximum.

prevailing species of Co^{2+} up to pH 8.0, and then the proportions of $Co(OH)^{+}$ and $Co(OH)_{2}$ become more important. When the pH value is higher than that for the formation of



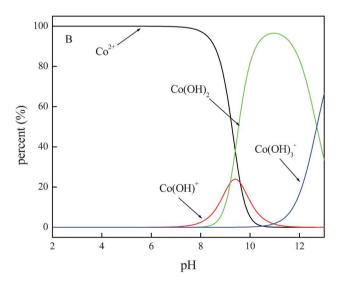


Fig. 5 Effect of pH on (A) the sorption of Co(II) to β-CD–GO and (B) the relative distribution of Co(II) species. A: T=303 K, m/V=0.1 g L $^{-1}$, I=0.01 M, NaCl, $C_{\text{[Co(II)]initial}}=10$ mg L $^{-1}$.

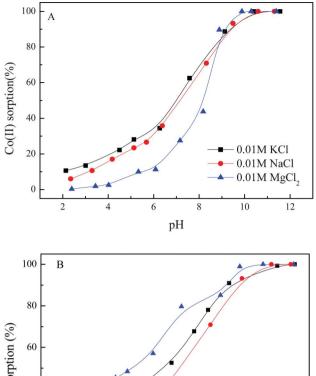
 $Co(OH)_2$, Co(II) is removed from the solution by sorption as well as by precipitation. At very low pH values, electrostatic interactions can hinder the sorption of Co(II) because both the Co(II) ions and the sorbent surface are positively charged, which explains the poor sorption of Co(II) ions on the β -CD–GO surface at low pH values. With increasing pH, the surface charge becomes more negative and the electrostatic interactions between Co(II) ions and β -CD–GO decrease, thereby resulting in an increase in Co(II) sorption.

The influence of ionic strength on the sorption of Co(II) onto β-CD-GO was investigated by a series of experiments using solutions containing different concentrations of NaCl. As shown in Fig. 5A, for all of the ionic strengths which were tested, Co(II) displays a weak affinity for β-CD-GO at low pH, suggesting that Co(II) does not easily form complexes with protonated carboxylic acid groups on the surface of β-CD-GO. No differences can be found at high pH under the experimental uncertainties, indicates that inner-sphere surface complexation is the main sorption mechanism. It is noteworthy that the sorption decreases significantly with increasing ionic strength at pH 6-9. The ionic strength dependence of Co(II) sorption onto β -CD-GO could be for two reasons: (1) metal ions form electrical double layer complexes with β-CD-GO, which favors a low ionic strength. This may indicate that the sorption interactions between the carboxylic acids on the surface of β-CD-GO and the Co(II) ions are partially ionic in nature; (2) the ionic strength of the solution influences the activity coefficient of the Co(II) ions, which limits their transfer to the surface of β-CD-GO.29

3.3. Effect of competing ions

To investigate the influence of competing ions on Co(II) sorption, the sorption of Co(II) on β-CD-GO was studied as a function of pH in 0.01 mol L⁻¹ NaCl, MgCl₂, KCl, NaNO₃ and NaClO₄ solutions, respectively. As can be seen from Fig. 6A, the sorption of Co(II) on β-CD-GO is strongly affected by the competing cations at pH < 9. The sorption of Co(II) on β -CD-GO under the same pH values is in the following sequence: K+ $> Na^+ > Mg^{2+}$, indicating that cations can alter the surface properties of β-CD-GO and thus can affect the sorption of Co(II) on β -CD-GO. The sorption of Co(II) on β -CD-GO can be considered as the competition of Co(II) with competing cations (K⁺, Na⁺, and Mg²⁺) at the β-CD-GO surface. The higher valence ion is much more readily sorbed onto the sorbent surface. Therefore, the influence of Mg²⁺ on Co(II) sorption is stronger than that of Na⁺ and K⁺. Because the radius of Na⁺ is smaller than that of K⁺, Na⁺ has a higher affinity to the surface of β-CD-GO. The higher tendency for counterion complexation with the surface groups of β -CD-GO reduces the number of ion interaction sites on the surface of β-CD-GO for Co(II) to interact with. In this work, the concentration of the competing cations $(1.00 \times 10^{-2} \text{ mol L}^{-1})$ is much higher than that of Co(II) (1.7 × 10⁻⁴ mol L⁻¹), and the influence of alkali metal cations and alkaline earth cations on the sorption of bivalent Co(II) is obvious. At pH < 9, the sorption of Co(II) on β -CD-GO can be attributed to the exchange of Co(II) with alkali metal cations or alkaline earth cations. It is reasonable that the coexisting competing cations can affect the sorption of Co(II) on β -CD-GO. One can also see that the sorption of Co(II) on

RSC Advances Paper



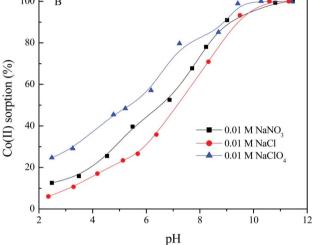


Fig. 6 Effect of ionic strength on the sorption of Co(II) to β-CD-GO. T = 303 K, $m/V = 0.1 \text{ g L}^{-1}$, $C_{[Co(II)]initial} = 10 \text{ mg L}^{-1}$. A: effect of cations, B: effect of anions.

 β -CD-GO at pH > 9 is not affected by competing cations, which may be due to the formation of surface precipitates at high pH values. 26,30

From Fig. 6B, one can see that the sorption of Co(II) on β-CD-GO is highest in 0.01 mol L⁻¹ NaClO₄ solution and is the lowest in 0.01 mol L^{-1} NaCl solution at pH < 9. Cl and NO₃ can form soluble complexes with Co(II), whereas ClO₄ cannot. Moreover, the acid radical radius order is Cl⁻ < NO₃⁻ < ClO₄⁻. The smaller radicals take up more ion exchange sites which leads to a decrease in Co(II) sorption.³¹

3.4. Effect of solid content

The effect of solid content on the sorption of Co(II) on β-CD-GO is shown in Fig. 7. The sorption of Co(II) increases with increasing solid content. This trend is expected because the number of functional sites which participate in the sorption of Co(II) increase with an increasing solid content. The distribution coefficient (K_d) decreases a little with an increasing solid content (Fig. 7). The decrease of K_d with an increasing solid content may be attributed to the competition with β -CD-GO.

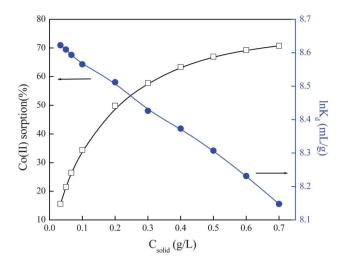


Fig. 7 The effect of solid content on the sorption of Co(II) to β -CD-GO. T = 303K, $C_{\text{[Co(II)]initial}} = 10 \text{ mg L}^{-1}$, I = 0.01 M NaCl, pH = 6.0 \pm 0.1.

Such competition reduces the effective functional group sites at the surface of β -CD-GO and results in a decrease of the sorption and complexation ability of β-CD-GO with an increasing solid content.

3.5. Sorption isotherms

The sorption isotherms obtained at 303, 323 and 343 K are shown in Fig. 8. With an increase in Co(II) concentration, the sorption initially increases rapidly and then increases more slowly. The sorption isotherms increase in the order 303, 323, and 343 K. This means that the process of sorption is favored at high temperatures. To gain a better understanding of the sorption mechanism and to quantify the sorption data, the

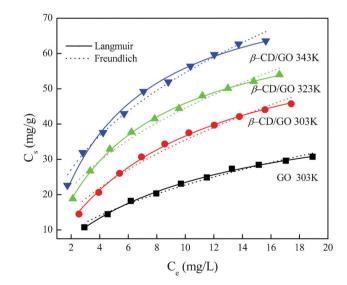


Fig. 8 Sorption isotherms of Co(II) on GO and β -CD-GO at three different temperatures. The solid line represents the Langmuir model and the dashed line represents the Freundlich model. pH = 6.0 \pm 0.1, m/V = 0.1 g L⁻¹, I = 0.01 M NaCl.

RSC Advances Paper

Langmuir and Freundlich models were used to simulate the experimental data (Fig. 9).

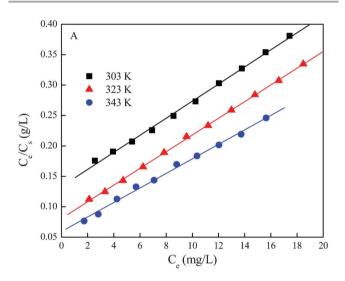
The Langmuir isotherm model was applied to describe the monolayer sorption process. It can be expressed by the following equation:

$$C_{\rm s} = \frac{bC_{\rm s\,max}C_{\rm e}}{1 + bC_{\rm e}}.\tag{3}$$

Eqn (3) can be expressed in the linear form:

$$\frac{C_{\rm e}}{C_{\rm s}} = \frac{1}{bC_{\rm s\,max}} + \frac{C_{\rm e}}{C_{\rm s\,max}} \tag{4}$$

where C_e is the equilibrium concentration of Co(II) remaining in solution (mg g^{-1}), $C_{s \text{ max}}$ (mg g^{-1}), the maximum sorption



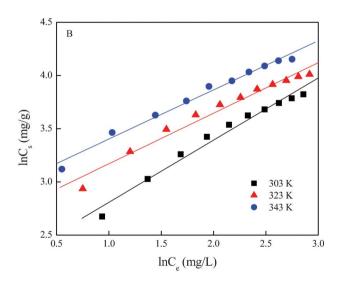


Fig. 9 Langmuir (A) and Freundlich (B) isotherms of the sorption of Co(II) on β-CD-GO at three different temperatures. pH = 6.0 \pm 0.1, m/V = 0.1 g L⁻¹, I = 0.01 M NaCl.

capacity, represents the amount of Co(II) adsorbed at a complete monolayer coverage, and b (L mg⁻¹) is a Langmuir constant which is related to the sorption heat.

The Freundlich isotherm model is applicable to several kinds of sorption on solid surfaces, which fits the sorption data at low and medium concentrations on heterogeneous surfaces. The model has the following form:

$$C_{\rm s} = K_{\rm F} C_{\rm e}^{\ n} \tag{5}$$

Eqn (5) can be expressed in linear form:

$$\ln C_{\rm s} = \ln K_{\rm F} + n \ln C_{\rm e} \tag{6}$$

where $K_{\rm F}$ (mg¹⁻ⁿ Lⁿ g⁻¹) represents the sorption capacity when the equilibrium concentration of the metal ions is equal to 1, and n represents the degree of the dependence of sorption with the equilibrium concentration.

Thermodynamic parameters were helpful for practical application. In order to investigate the impact of temperature, the sorption isotherms of β -CD/GO at different temperatures were obtained and shown in Fig. 8. The sorption of Co(II) increases with an increase in the reaction temperature. The highest sorption isotherm is obtained at T = 343 K, while the lowest sorption isotherm is obtained at T = 303 K, which illustrates that the sorption of Co(II) on β-CD-GO is promoted at higher temperatures. The thermodynamic parameters (ΔG^0 , ΔS^0 and ΔH^0) are calculated from the temperature-dependent sorption isotherms. The value of the free energy change is derived from the function:

$$\Delta G^0 = -RT \ln K^0 \tag{7}$$

where K^0 is the sorption equilibrium constant. The values of ln K^0 represent the sorption ability of β-CD-GO towards Co(II). K^0 can be calculated by plotting ln K_d versus C_e and extrapolating $C_{\rm e}$ to zero (see Fig. 10). The standard enthalpy change (ΔH^0)

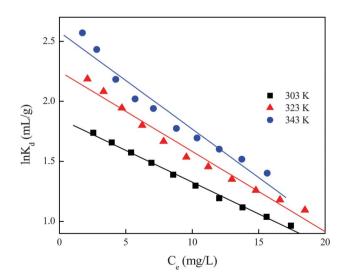


Fig. 10 Linear plots of $\ln K_d$ versus C_e . pH = 6.0 \pm 0.1, m/V = 0.1 g L⁻¹, I = 0.01 M NaCl.

and the standard entropy change (ΔS^0) are calculated from the following equation:

$$\ln K^0 = \frac{\Delta H^0}{R} - \frac{\Delta S^0}{RT} \tag{8}$$

Linear plots of $\ln K^0$ vs. 1/T for Co(II) sorption on β-CD-GO are shown in Fig. 11. R (8.314 J mol⁻¹ K⁻¹) is the ideal gas constant and T (K) is the temperature.

The change of Gibbs free energy (ΔG) was calculated from the equation:

$$\Delta G^0 = \Delta H^0 - T \Delta S^0. \tag{9}$$

The values calculated from the thermodynamic equations are listed in Table 3. The positive value of ΔH^0 indicates that the sorption is endothermic, which is in accordance with the fact that the sorption increases as the reaction temperature increases. A probable interpretation for the positive enthalpy is that the Co(II) ions are completely solvated in water, the hydration sheath of Co(II) is supposed to be destroyed before the ions adsorb onto the β -CD–GO surface, and the energy required for this dehydration process exceeds the energy released when the cations are attached to the solid surface, 32 meaning a high temperature is favored. The negative value of the change in Gibbs free energy (ΔG^0) indicates that the sorption is a spontaneous process under the applied experimental conditions. 33

As the reaction temperature increases, the values of ΔG^0 become even more negative, which illustrates that sorption is more efficient at higher temperature. At high temperature, cations are likely to be more desolvated, which favors the sorption reaction.³⁴ The positive value of the entropy change (ΔS^0) may be attributed to the structure change of the composite complex.³⁵ It can be concluded that the sorption of Co(II) on β -CD-GO is an endothermic and spontaneous

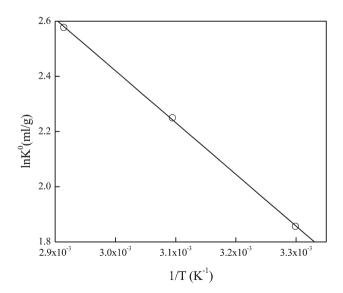


Fig. 11 Linear plots of ln K^0 vs. 1/T for Co(II) sorption on β -CD-GO.

Table 3 Thermodynamic parameters of Co(II) sorption on β-CD-GO

T(K)	ΔG^0 (kJ mol ⁻¹)	ΔS^0 (J K $^{-1}$ mol $^{-1}$)	ΔH^0 (kJ mol ⁻¹)	
303	-4.68	66.93	15.61	
323	-6.04	66.93	15.58	
343	-7.35	66.93	15.61	

process according to the thermodynamic data of the sorption isotherms.

From the fitting curves and the correlation coefficients listed in Table 4, it is clear that the Langmuir model simulates the experimental data better than the Freundlich model, which indicates that monolayer sorption occurs and the interaction between the ions is weak and can be neglected.36 The $C_{\rm s\ max}$ of Co(II) sorption on β-CD-GO at pH 6.0 \pm 0.1 and T = 303 K was about 72.43 mg g⁻¹, which is much higher than that of Co(II) on GO (47.39 mg g^{-1}). This means that under the same experimental uncertainties, in comparison to GO, the modified β-cyclodextrins on GO can reduce the amount of sorbent that is needed by \sim 35 wt%. Moreover, in comparison to the $C_{\rm s~max}$ values of Co(II) sorption on common sorbents, such as Al-pillared bentonite (38.6 mg g⁻¹ Co(II) at pH 6 and T= 303 K),³⁷ lemon peel (25.6 mg g⁻¹ Co(II) at pH 6 and T = 298 K),³⁸ activated carbon (1.2 mg g⁻¹ Co(II) at pH 6 and T = 303K), 39 and magnetite-GO (12.98 mg g⁻¹ Co(II) at pH 6.8 and T =303 K), 40 one can see that β-CD-GO has a much higher sorption capacity than other commonly used sorbents. It indicates that β-CD-GO may be a high-efficiency material in the clean-up of heavy metal ion-based pollution.

4 Conclusions

The results of SEM, FT-IR, Raman and XPS analysis indicated that β -CD was successfully modified onto the surface of GO. The ionic strength-independent and pH-dependent Co(II) sorption on β-CD-GO indicated that the sorption mechanism of Co(II) featured outer-sphere surface complexation at low pH values, whereas the removal of Co(II) was accomplished by simultaneous precipitation and inner-sphere surface complexation at high pH values. The sorption of Co(II) on β-CD-GO was also influenced by competing ions. The thermodynamic parameters indicated that the interaction of Co(II) with β-CD-GO was an endothermic and spontaneous process. The main sorption mechanism for Co(II) adsorbed onto β-CD-GO was the surface complexation of Co(II) with abundant oxygencontaining functional groups on the surface of β-CD-GO. The maximum sorption capacities ($C_{s \text{ max}}$) of Co(II) on β -CD-GO at pH 6.0 \pm 0.1 and T = 303 K was about 72.43 mg g⁻¹, which is much higher than that of Co(II) on GO. β-CD-GO exhibits the potential for application as a sorbent in the clean-up of environmental pollution.

Table 4 Parameters of the Langmuir and Freundlich isotherms at different temperatures

Sorbent	T(K)	Langmuir		Freundlich			
		$C_{\rm s~max}~({\rm mg~g}^{-1})$	b (L mg ⁻¹)	R^2	$k_{\rm F} \left({\rm mg}^{1-n} \; {\rm L}^n {\rm g}^{-1} \right)$	n	$R^2 \left(mg^{1-n} L^n g^{-1} \right)$
GO	303	47.39	0.983	0.998	6.28	0.562	0.988
β-CD-GO	303	72.43	0.1051	0.997	9.24	0.585	0.977
•	323	72.99	0.1696	0.999	14.81	0.475	0.971
	343	84.03	0.2010	0.996	18.95	0.462	0.985

Acknowledgements

Financial support from the National Natural Science Foundation of China (21007074, 21107115, 11205203, 21207136) is acknowledged.

References

- 1 C. L. Chen, D. Xu, X. L. Tan and X. K. Wang, *J. Radioanal. Nucl. Chem.*, 2007, 273, 227–233.
- 2 S. Rengaraj and S. H. Moon, Water Res., 2002, 36, 1783-1793.
- 3 S. I. El-Dessouky, E. A. El-Sofany and J. A. Daoud, *J. Hazard. Mater.*, 2007, **143**, 17–23.
- 4 D. Xu, D. D. Shao, C. L. Chen, A. P. Ren and X. K. Wang, *Radiochim. Acta*, 2006, **94**, 97–102.
- 5 O. Yavuz, Y. Altunkaynak and F. Guzel, *Water Res.*, 2003, 37, 948–952.
- 6 X. Ren, C. Chen, M. Nagatsu and X. Wang, *Chem. Eng. J.*, 2011, 170, 395–410.
- 7 S. Yang, J. Hu, C. Chen, D. Shao and X. Wang, *Environ. Sci. Technol.*, 2011, 45, 3621–3627.
- 8 E. D. van Hullebusch, J. Gieteling, M. Zhang, M. H. Zandvoort, W. Van Daele, J. Defrancq and P. N. L. Lens, J. Biotechnol., 2006, 121, 227–240.
- 9 A. F. Ngomsik, A. Bee, D. Talbot and G. Cote, Sep. Purif. Technol., 2012, 86, 1–8.
- 10 J. Snukiškis and D. Kaušpėdienė, *Colloids Surf.*, A, 2005, 253, 27–32.
- 11 P. M. van Berkel, M. Punt, G. J. A. A. Koolhaas, W. L. Driessen, J. Reedijk and D. C. Sherrington, *React. Funct. Polym.*, 1997, 32, 139–151.
- 12 Y. W. Zhu, S. Murali, W. W. Cai, X. S. Li, J. W. Suk, J. R. Potts and R. S. Ruoff, Adv. Mater., 2010, 22, 3906–3924.
- 13 G. Zhao, X. Ren, X. Gao, X. Tan, J. Li, C. Chen, Y. Huang and X. Wang, *Dalton Trans.*, 2011, **40**, 10945–40952.
- 14 G. Zhao, J. Li and X. Wang, *Chem. Eng. J.*, 2011, 173, 185–190.
- 15 G. Zhao, T. Wen, C. Chen and X. Wang, *RSC Adv.*, 2012, 2, 9286–9303.
- 16 G. X. Zhao, L. Jiang, Y. D. He, J. X. Li, H. l. Dong, X. k. Wang and W. P. Hu, Adv. Mater., 2011, 23, 3959–3963.
- 17 M. Arkas, R. Allabashi, D. Tsiourvas, E. M. Mattausch and R. Perfler, *Environ. Sci. Technol.*, 2006, **40**, 2771–2777.
- 18 W. S. Mnger, J. Jacob, K. Gessler, T. Steiner and D. Hoffmann, Chem. Rev., 1998, 98, 1787–1802.

- 19 E. C. Galia, O. Navarre, M. Scialdone, G. Ferreira, S. Filardo and E. Tilloy, *J. Phys. Chem. B*, 2007, **111**, 2573–2578.
- S. Berto, M. C. Bruzzoniti, R. Cavalli, D. Perrachon,
 E. Prenesti, C. Sarzanini, F. Trotta and W. Tumiatti, J. Inclusion Phenom. Macrocyclic Chem., 2007, 57, 631–636.
- 21 M. Hirata, T. Gotou, S. Horiuchi, M. Fujiwara and M. Ohba, *Carbon*, 2004, **42**, 2929–293.
- 22 Y. Sun, Q. Wang, C. Chen, X. Tan and X. Wang, *Environ. Sci. Technol.*, 2012, **46**, 6020–6027.
- 23 Y. J. Guo, S. J. Guo, J. Li, E. K. Wang and S. J. Dong, *Talanta*, 2011, **84**, 60–64.
- 24 J. Hu, D. D. Shao, C. L. Chen and J. X. Li, *J. Phys. Chem. B*, 2010, **114**, 6779–6785.
- 25 D. D. Shao, G. D. Sheng, C. L. Chen and X. K. Wang. M. Nagatsu, *Chemosphere*, 2010, **79**, 679–685.
- 26 Q. Wang, J. X. Li, C. L. Chen and X. M. Ren, *Chem. Eng. J.*, 2011, 174, 126–133.
- 27 G. Zhao, T. Wen, X. Yang, S. Yang, J. Liao, J. Hu, D. Shao and X. Wang, *Dalton Trans.*, 2012, **41**, 6182–6188.
- 28 H. Yüzer, M. Kara, E. C. Sabah and M. S. Celik, *J. Hazard. Mater.*, 2008, **151**, 33–37.
- 29 Z. Reddad, C. Gerente, Y. Andres and L. P. Cloirec, *Environ. Sci. Technol.*, 2002, 36, 2067–2073.
- 30 X. L. Tan, X. K. Wang, H. Geckeis and T. H. Rabung, Environ. Sci. Technol., 2008, 42, 6532–6537.
- 31 S. T. Yang, J. X. Li, D. D. Shao, J. Hu and X. K. Wang, *J. Hazard. Mater.*, 2009, **166**, 109–116.
- 32 G. D. Sheng, J. X. Li, D. D. Shao, J. Hu, C. L. Chen, Y. X. Chen and X. K. Wang, *J. Hazard. Mater.*, 2010, **178**, 333–340.
- 33 J. Li, S. Zhang, C. Chen, X. Yang, J. Li and X. Wang, ACS Appl. Mater. Interfaces, 2012, 4, 4991–5000.
- 34 S. S. Tahir and N. Rauf, J. Chem. Thermodyn., 2003, 35, 2003–2009.
- 35 H. Genc-Fuhrman, J. C. Tjell and D. Mcconchie, *Environ. Sci. Technol.*, 2004, **38**, 2428–2434.
- 36 J. Hu, D. L. Zhao and X. K. Wang, Water Sci. Technol., 2011, 63, 917–923.
- 37 D. M. Marlohar, B. F. Noeline and T. S. Anirudhan, *Appl. Clay Sci.*, 2006, 31, 194–206.
- 38 A. Bhatnagar, A. K. Minocha and M. Sillanpaa, *Biochem. Eng. J.*, 2010, **48**, 181–186.
- 39 H. Subaymon, B. A. Abid and J. A. Al-Najar, *Chem. Eng. J.*, 2009, 155, 647–653.
- 40 M. C. Liu, C. L. Chen, J. Hu, X. L. Wu and X. K. Wang, *J. Phys. Chem. C*, 2011, **115**, 25234–25240.